## ENVIRONMENTAL RESEARCH LETTERS

## PAPER • OPEN ACCESS

# Accounting for re-exports substantially reduces China's virtual water demand through agricultural trade

To cite this article: Tariq Ali et al 2021 Environ. Res. Lett. 16 045002

View the article online for updates and enhancements.

# ENVIRONMENTAL RESEARCH

## CrossMark

LETTER

**OPEN ACCESS** 

RECEIVED 1 December 2020

REVISED

21 February 2021

25 February 2021

PUBLISHED 12 March 2021

Original content from this work may be used under the terms of the Creative Commons Attribution 4.0 licence.

Any further distribution of this work must maintain attribution to the author(s) and the title of the work, journal citation and DOI.



# Accounting for re-exports substantially reduces China's virtual water demand through agricultural trade

#### Tariq Ali<sup>1</sup>, Wei Xie<sup>2,\*</sup>, Anfeng Zhu<sup>2,3</sup> and Kyle Frankel Davis<sup>4,5</sup>

<sup>1</sup> School of Economics and Management, Jiangxi Agricultural University, Nanchang 330045, People's Republic of China
 <sup>2</sup> China Center for Agricultural Policy, School of Advanced Agricultural Sciences, Peking University, No. 5 Yiheyuan Road, Haidian

District, Beijing 100871, People's Republic of China

- School of Economics and Management, North China University of Technology, No. 5 Jinyuanzhuang Road, Shijingshan District, Beijing 100144, People's Republic of China
- Department of Geography and Spatial Sciences, University of Delaware, Newark, DE 19716, United States of America
- Department of Plant and Soil Sciences, University of Delaware, Newark, DE 19716, United States of America

\* Author to whom any correspondence should be addressed.

E-mail: xiewei.ccap@pku.edu.cn

Keywords: global value chains, virtual water trade, agriculture

Supplementary material for this article is available online

#### Abstract

Traditional methods of assessing virtual water (VW) trade usually overestimate the foreign VW used by a country as the imported VW can also be re-exported and used by the countries other than the initial importer. Due to its ability to provide better estimates of transboundary resource use, the global value chain (GVC) method can provide more precise guidelines for the global policy debate over sustainable resource use. Here, we use GVC analysis based on multi-regional input-output tables to quantify the embodied trade of virtual blue and green water for China in the major agricultural sectors and agro-based industries. We find that China is a net importer of blue VW (2.9 billion m<sup>3</sup>) and green VW (57.9 billion m<sup>3</sup>) through agricultural commodities. Our results reveal that a large portion of imported blue (37%) and green (17%) VW is re-exported by China and consumed in other countries, representing the overestimated parts of China's VW import use. These ratios are even higher for individual commodities, including 44% for blue VW for cotton and 22% for green VW for grains. This work demonstrates the importance of improved accounting in VW trade to achieve the sustainable use of global water resources and equitably share responsibility between producers, intermediates, and end-users.

### 1. Introduction

A large body of literature has analyzed the transboundary movement of embodied resources through trade in goods and services (Konar *et al* 2012, Lin *et al* 2014, Meier *et al* 2014, Dalin *et al* 2015, Vora *et al* 2017, Zhang *et al* 2017a, Zhao *et al* 2018, Deng *et al* 2020). Due to its scarcity (Mekonnen and Hoekstra 2016, Veldkamp *et al* 2017) and importance for societies through food and agricultural systems (Davis *et al* 2017, D'Odorico *et al* 2018), economies (UNESCO 2016) and ecosystem services (Grizzetti *et al* 2016), water traded as virtual water (VW)<sup>6</sup> has been widely studied (Fader *et al* 2011, Dalin *et al* 

<sup>6</sup> The water embodied in traded commodities (Allan 1997).

2012, Orlowsky et al 2014, Vörösmarty et al 2015, Flach et al 2016, Salmoral and Yan 2018, Qian et al 2019, Zhao et al 2019). Agriculture, the largest user of freshwater, has been the main focus of this literature. The traditional 'bottom-up' approach is most commonly used in these studies, employing detailed process data mostly for agricultural products (Chapagain et al 2006, Zhang et al 2016, Ali et al 2017). Despite its utility, this approach fails to account for inputs used at various stages of production. Measuring the trade in gross values through the bottom-up approach, where we hold the last country in the supply chain accountable for the entire value of export, potentially provides an inaccurate picture of international trade relations and may wrongly attribute production and consumption to various geographic locations (Elms and Low 2013).

A 'top-down' approach based on the concept of global value chains (GVCs) can eliminate the estimation errors within the 'bottom-up' approach (Koopman et al 2014, Elms and Low 2013). GVC is based on the concept that goods produced by firms in one part of the world can then be used as inputs by firms in their production process in other places (WTO 2019). Intermediate inputs move between various destinations multiple times; therefore, traditional trade statistics are becoming less reliable in gauging the value contributed by any particular country (Koopman et al 2014). In GVC terms, forwardlinkages describe a country's action to sell its exports as inputs for other countries' exports, and backward linkages describe a country's action when it buys inputs from other countries for its exports (Greenville et al 2019). GVC can thus provide a clearer picture of VW flows associated with imports and exports along the entire global supply chain while maintaining a full system boundary (Feng et al 2011)<sup>7</sup>.

China occupies a unique position in global VW trade through agricultural commodities due to its expanding agricultural trade, water scarcity, and increasing role in forward- and backward-linkages of GVCs of agriculture. With the world's lowest per capita water availability (FAO 2016), China's domestic water resources are under immense pressure. In addition to several other factors, China's reliance on foreign agricultural products has been partially driven by the increased stress on its domestic water resources. For example, since joining the World Trade Organization, the country's total food imports and exports increased from US\$ 15 billion to US\$ 127 billion and US\$ 24 billion to US\$ 92 billion during 2002-2017, respectively. The expanded trade in agro-based commodities was also profound, with, for instance, China's exports and imports of textiles growing from US\$ 62 billion and US\$ 12 billion in 2002 to US\$ 298 billion and US\$ 24 billion in 2017 (FAOSTAT 2021). Moreover, during 2004-2014, China contributed the highest share (21%) in the global growth of forwardlinkages in GVCs of agriculture and food trade (Greenville et al 2019).

These factors call for a deeper understanding of the consequences of China's expansion in GVC networks and its global impact on water and other natural resources. Several existing studies show that China's trade in VW through agricultural products has been growing (Chapagain and Hoekstra 2004, Liu *et al* 2007, Dalin *et al* 2014, Shi *et al* 2014, Zhuo *et al* 2016, Ali *et al* 2017). Although using different commodity mixes and time scales, most of these studies have employed 'bottom-up' analyses to show that China is a net importer of VW through agricultural commodity trade. For example, China's net-imports of VW through agriculture trade were reported as 31 billion m<sup>3</sup> yr<sup>-1</sup> during 1991-2004 (Liu et al 2007); 138 billion m<sup>3</sup> in 2009 (Shi et al 2014); and 120 billion m<sup>3</sup> in 2015 (Ali et al 2017). However, these analyses may present an incomplete picture of China's trade in VW through agricultural commodities as they assume that China consumes all the imported VW. While certain studies on VW trade have attempted to use the input-output approach to overcome the overestimation problem of VW imports by China, they have either only covered China's interprovincial trade using China's input/output table or had a coarse representation of agriculture as a single sector (e.g. Zhao et al 2010, Wang et al 2016, Yang et al 2016, Hou et al 2018). Therefore, a comprehensive analysis is still needed to assess China's import, export, and re-export of blue and green VW internationally.

In this study, we construct a multi-region inputoutput (MRIO) model, integrating direct resource use and monetary flows based on a global economic model for 2011-the latest year for which the data was available. We combine the MRIO model with VW contents (blue and green) of eight major crops to analyze the intermediate and final demand for VW by consumers in China and other countries through primary and processed agricultural commodities. The ultimate objective is to rectify the overestimation problem inherent in the traditional estimation methods by using a more suitable approach for the trade of VW through China's trade in agricultural commodities to various regions. Using such an approach can lead to more accurate estimates of nations' virtual resource imports and exports globally. These findings can provide vital information for improving international coordination between China and its trade partners on achieving the sustainable use of freshwater resources.

## 2. Methods and data

We used a MRIO model to analyze the VW use associated with China's agricultural trade, integrating direct resource use and monetary flows (see SI figure S2 (available online at stacks.iop.org/ERL/16/045002/mmedia) for MRIO schematic table).

# 2.1. Measuring embodied water use by GVC decomposition method

We assume a world of G countries and N sectors. These countries are connected through the inter-regional trade of intermediate and final products, and each country's outputs are used to satisfy intermediate or final demand by domestic and foreign consumers, such that:

<sup>&</sup>lt;sup>7</sup> For a detailed description and comparison of 'bottom-up' and 'top-down' approaches see Feng *et al* (2011).

$$\begin{bmatrix} X^{1} \\ X^{2} \\ \vdots \\ X^{g} \end{bmatrix} = \begin{bmatrix} A^{11} & A^{12} & \cdots & A^{1g} \\ A^{21} & A^{22} & \cdots & A^{2g} \\ \vdots & \vdots & \ddots & \vdots \\ A^{g1} & A^{g2} & \cdots & A^{gg} \end{bmatrix} \begin{bmatrix} X^{1} \\ X^{2} \\ \vdots \\ X^{g} \end{bmatrix} + \begin{bmatrix} \sum_{r}^{G} Y^{1r} \\ \sum_{r}^{G} Y^{2r} \\ \vdots \\ \sum_{r}^{G} Y^{gr} \end{bmatrix}$$
(1)

where  $X^{s}$  represents the gross output of country s (s = 1, ..., g),  $Y^{sr}$  represents the final demand of country r (r = 1, ..., g) for products from country s,  $A^{sr}$  is the input coefficient matrix representing the intermediate use in country r of goods produced

in country *s*. The elements of the input coefficient matrix satisfy  $a_{ij}^{sr} = z_{ij}^{sr}/x_j^r$ , where  $z_{ij}^{sr}$  (*i*, *j* = 1, ..., *n*) represents the transfer from sector *i* of country *s* to sector *j* of country *r*. The intermediate input matrix from country *s* to country *r* is represented by  $z^{sr} = A^{sr}X^r$ . Equation (1) can be rearranged as

$$\begin{bmatrix} X^{1} \\ X^{2} \\ \vdots \\ X^{g} \end{bmatrix} = \begin{bmatrix} I - A^{11} & -A^{12} & \cdots & -A^{1g} \\ -A^{21} & I - A^{22} & \cdots & -A^{2g} \\ \vdots & \vdots & \ddots & \vdots \\ -A^{g1} & -A^{g2} & \cdots & I - A^{gg} \end{bmatrix}^{-1} \begin{bmatrix} \sum_{r}^{G} Y^{1r} \\ \sum_{r}^{G} Y^{2r} \\ \vdots \\ \sum_{r}^{G} Y^{gr} \end{bmatrix} = \begin{bmatrix} B^{11} & B^{12} & \cdots & B^{1g} \\ B^{21} & B^{22} & \cdots & B^{2g} \\ \vdots & \vdots & \ddots & \vdots \\ B^{g1} & B^{g2} & \cdots & B^{gg} \end{bmatrix}^{-1} \begin{bmatrix} \sum_{r}^{G} Y^{1r} \\ \sum_{r}^{G} Y^{2r} \\ \vdots \\ \sum_{r}^{G} Y^{gr} \end{bmatrix}$$
(2)

where  $B^{sr}$ , the Leontief inverse, represents the quantity of the gross output of country *s* for a one-unit increase in the final demand of country *r*. From equation (2), the gross output of country *r* can be given as:

The intermediate input of country *r* from country *s* is  $Z^{sr} = A^{sr}X^r$ . The exports from country *s* to country *r* are  $T^{sr} = Y^{sr} + A^{sr}X^r$  and can be divided into three parts (Wang *et al* 2013) as

$$X^{r} = \sum_{G}^{t} B^{rt} \sum_{G}^{u} Y^{tu}.$$
 (3)

 $T^{sr} = Y^{sr} + A^{sr}X^{r}$   $= Y^{sr} + A^{sr}B^{rr}Y^{rr} + A^{sr}B^{rs}Y^{sr} + A^{sr}\sum_{\substack{t \neq s, r}}^{G}B^{rt}Y^{tr} + A^{sr}B^{rr}Y^{rs} + A^{sr}B^{rs}Y^{ss} + A^{sr}\sum_{\substack{t \neq s, r}}^{G}B^{rt}Y^{ts}$   $+ A^{sr}B^{rr}\sum_{\substack{t \neq s, r}}^{G}Y^{rt} + A^{sr}B^{rs}\sum_{\substack{t \neq s, r}}^{G}Y^{st} + A^{sr}\sum_{\substack{t \neq s, r}}^{G}B^{rt}\sum_{\substack{u \neq s, r}}^{G}Y^{tu}.$  (4)

 $W^{ps} = F^{ps} X^s. ag{5}$ 

We define sectoral water intensity as  $f^{ps} = w^{ps}/x^{ps}$ , where  $w^{ps}$  represents the water requirement of an agricultural sector p of country s.  $F_{ps} = (f_{p_1s}, f_{p_2s}, \ldots, f_{p_xs}, 0, \ldots, 0)'$  is a diagonal matrix composed of  $f_{ps}$ . The water consumption of country s is

W<sup>ps</sup> is the water consumed in the country's agricultural sectors, which we can divide into two parts based on the form it is consumed. The first part is the water embodied in agricultural products that are consumed as final products. The second part is the water embodied in the production that is consumed as intermediate goods. The water embodied in intermediate goods is

$$W^{nps} = \left[ \left( W^{ps} - F^{ps} \operatorname{diag} \left( A^{ss} X^{s} \right) - Y^{ss} - \sum_{r \neq s}^{G} T^{sr} \right) C^{ss} \right]'$$
(6)

where  $C^{ss} = (Z^{ss} - \text{diag}(Z^{ss})) \times (X^s))^{-1}$ , it presents the ratio of the agricultural products consumed in non-agricultural sectors, and the water consumption intensity of the *np* sector in *s* country is  $f^{nps} = W^{nps}/(Y^{ss} + T^{sr})$ . Therefore, the water consumption intensity of all sectors in the country *s* is

$$F^{s} = (f^{p_{1}s}, f^{p_{2}s}, \dots, f^{p_{x}s}, f^{np_{1}s}, f^{np_{2}s}, \dots, f^{np_{y}s})'.$$
(7)

 $F^s$  represents the increase in domestic water caused by an increase of one unit in the country *s* exports, agricultural sectors x (x = 1, ..., x), and non-agricultural sectors y (y = 1, ..., y). Thus, the domestic water embodied in the exports from country *s* to country *r* in equation (6) can be given as:

$$WEX^{sr} = F^{s}T^{sr} f^{r} + F^{s}T^{sr} f^{s} + F^{s}T^{sr} f^{other}.$$
 (8)

Based on equation (8), the gross domestic water of country s exported to country r is decomposed into three paths. Path 1  $(F^{s}T^{sr}_{f})$  shows the water embodied in products exported from country s to country r, which are directly consumed or used as intermediate products for production in other sectors by country r. Path 2  $(F^{s}T^{sr}_{f})$  is the water embodied in the commodities exported from country s to country r, which are used by country r to produce other commodities and then returned and finally consumed by country s. Path 3 ( $F^{s}T^{sr}_{f}f^{other}$ ) shows the water embodied in the products exported from country s to country r, which are used by country r to produce other commodities and then reexported to other countries (excluding countries s and *r*) and finally consumed by other countries. Each part of equation (8) represents the domestic water embodied in the products produced in the country s.

#### 2.2. Data

#### 2.2.1. Monitory MRIO table

This study uses the latest version of the Global Trade Analysis Project-Water (GTAP-W) database (2011 base-year), which contains 140 countries/regions and 76 sectors (Haqiqi *et al* 2016). We aggregate the GTAP-W database into 18 sectors while ensuring that all the primary and processed agricultural sectors are present in the most disaggregated form. We also aggregate the GTAP countries into 12 regions while keeping the detailed representation of China's main trading partners (SI tables S1 and S2 contain details of regional and sectoral aggregations) (available online at stacks.iop.org/ERL/16/045002/mmedia).

#### 2.2.2. Agricultural blue water (irrigation) use

We used the following method to estimate the sectoral water intensity ( $f^{ps} = w^{ps}/x^{ps}$ ) for blue VW for all

the sectors and regions in our aggregated database, by dividing the total water requirement of agricultural sector p of country s ( $w^{ps}$ ) with the gross output of the sector p in country s ( $x^{ps}$ ).

- (a) Crop water requirement (CWR): this data set is based on the work by Siebert and Döll (2010) and represents worldwide information on water requirements for irrigation for 29 crop categories at a 5 min spatial resolution in 2000. In our study, we assume that the water requirements in 2011 are similar to those in 2000<sup>8</sup>.
- (b) Irrigation water requirement (IWR) by country in 2011: this data set is obtained from the FAO's global water information system (2016). We then use the CWR for each crop to allocate each country's IWR to each crop sector in our aggregated database.

#### *2.2.3.* Agricultural use of green water (soil moisture)

Due to data limitations, we rely on the blue water use estimated in the previous subsection and ratios between the blue and green water requirements for crop production to obtain the green water use by crops based on the blue and green water use per unit of the crop, as described in Hanasaki (2016). We then get the green water use through the following equation:

$$GW_{i,r} = (G/B) \operatorname{Ratio}_{i,r} \times BW_{i,r}$$
 (9)

where  $GW_{i,r}$  is the amount of green water embodied in the production of per unit of a crop *i* in country/region *r*;  $\frac{G}{B}$ Ratio<sub>*i*,*r*</sub> is the ratio of green to blue water and  $BW_{i,r}$  is the blue water used in per unit production of crop *i* and country/region *r*.

#### 3. Results

Here we present the main net import results of our GVC analysis of China's VW trade associated with primary and processed agricultural commodities in 2011.

# 3.1. Net VW imports by China through different commodities

The main commodities for total agricultural VW import are soybean (80.9 billion m<sup>3</sup>), cotton (32.0 billion m<sup>3</sup>), and maize (3.0 billion m<sup>3</sup>) from the USA, Brazil, Argentina, India, and Central Asia. The main commodities for export of VW

<sup>8</sup> For a more accurate representation of water intensities for rice and wheat, we used VWC from Hanasaki (2016) and multiplied them with the total quantity of rice and wheat output in each country to get total water requirement for them. We then subtract the water requirement for rice and wheat from the national IWR. For the rest of the crops, we divided the remaining IWR by using CWR for these crops from Siebert and Döll (2010).



are 'other crops' (including crops like coffee, tea, and spices), processed food, and textile mainly to South Asia, Southeast Asia, European Union, and North America (figure 1(a); see supplementary figure S1 for imports and exports of blue and green VW embodied in primary and processed agricultural commodities).

China is a net importer of blue VW through primary agricultural commodities (15.4 billion  $m^3$ , the sum of the bars in the upper part of figure 1(b)) but a net exporter of blue VW through processed agricultural commodities (12.4 billion  $m^3$ , the sum of the bars in the lower part of figure 1(a)). The total effect is a net-import of 2.9 billion  $m^3$  of blue VW by China. The highest net import of blue VW is through cotton (17.8 billion  $m^3$ ) and oilseeds (3.1 billion  $m^3$ , which is mostly soybean). Sectors such as 'other crops' (including coffee, tea, and spices) are the main primary agricultural commodities responsible for blue VW's net-export, with net-exports of 3.6 billion m<sup>3</sup>. For the processed agricultural commodities, the net exports are dominated by textile (processed cotton) and leather (12.4 billion m<sup>3</sup>), followed by processed food sectors (0.1 billion m<sup>3</sup>).

For green VW, China is a net importer for most agricultural commodities, with a total net-import of 57.9 billion m<sup>3</sup> (figure 1(c)). In terms of the primary agricultural commodities, oilseeds (72.0 billion m<sup>3</sup>) and cotton (14.1 billion m<sup>3</sup>) dominate green VW's net-import. Chicken and pork (2.7 billion m<sup>3</sup>) and other crops (1.0 billion m<sup>3</sup>) are the top primary agricultural commodities responsible for green VW's highest net-exports from China. Among the processed agricultural commodities, textile and



leather (29.7 billion m<sup>3</sup>) and processed food sectors (1.3 billion m<sup>3</sup>) show considerable net export of green VW by China (figure 1(c)). The combined netimports of both blue and green VW through primary agricultural commodities (rice, wheat, other grains<sup>9</sup>, vegetables, fruits, oilseeds, beef, mutton, chicken, and pork) by China stood at 75.3 billion m<sup>3</sup>, comparable to earlier findings of 76.8 billion m<sup>3</sup> by Ali *et al* (2017) for the same period. Shi *et al* (2014) used the traditional 'bottom-up' approach and reported a net import of ~138 billion m<sup>3</sup> of VW by China through 27 major primary crops in 2009. Their estimates are higher as they covered more crops and did not report the re-export of VW.

# 3.2. Net VW imports by China through different pathways

We find that not all of the VW imported by China through agricultural commodities is ultimately consumed within China. Instead, large fractions of this VW are re-exported via primary and processed agricultural commodities, either back to the source country or onward to a third country (figure 2). Although China's combined (positive)

<sup>9</sup> As maize is the largest part of imports by China among all the grain crops aggregated in the 'Other grains' sector, the results for 'Other grains' closely depict China's VW trade in maize.

net-imports of blue and green VW through agricultural commodities were 114.2 billion  $m^3$ , the nation re-exported 24.1 billion  $m^3$  (or 21%) of this imported VW to other countries (sum of respective shares in figure 2(a)).

When disaggregated between blue and green VW, we find that China re-exports 37% of imported blue and 17% of imported green VW to other countries for their final consumptive use. For cotton, the biggest net-importing commodity of blue VW for China, we see that about 56% (9.9 billion  $m^3$ ) of blue VW imported by China is used for domestic consumption, while the rest, 7.9 billion m<sup>3</sup>, is re-exported to the outside world via different trade pathways. About 41% (7.3 billion m<sup>3</sup>) of blue VW net-import through cotton is re-exported to third countries, while 3% (0.6 billion m<sup>3</sup>) returns to source countries. For oilseeds (including 'soybean') and sugar crops, the ratios of domestic use of imported blue VW by China are substantially higher (88% and 96%, respectively), showing that most of the blue VW imported through oilseeds and sugar is consumed within China, and the rest (12% and 4%, respectively) is re-exported in the form of intermediate or final goods.

For green VW, we see that China uses a significant portion (88%) of the imported green VW (63.5 billion  $m^3$ ) through the imports of oilseeds for its domestic use, while 10% (7.2 billion  $m^3$ )



VW by China and then consumed in China, source, or third countries.

and 2% (1.1 billion  $m^3$ ) shares are re-exported to third countries and source countries, respectively (figure 2(b)). We observe that a significant portion (44%) of green VW imported by China through cotton trade is re-exported to other world regions. For wheat (14%), grains (which is mostly maize for China, 23%), and sugar crops (11%), reasonably large ratios of the green VW imported by China are reexported to other countries.

# 3.3. Sources and destinations of China's net VW imports

Our findings also show that China's imports of blue VW through cotton are mainly sourced from India (South Asia) and Central Asian countries (72%), followed by the USA (North America, 23%) (figure 3). Interestingly, 43% of the blue VW imported through China's cotton trade is re-exported to third countries, while only 56% is retained for its domestic use. About 11% of the blue VW imported through cotton trade by China from the USA (North America) is returned to the USA, while 33% is re-exported to other countries. China imports about 97% of its blue VW through oilseeds (soybean) from Brazil, Argentina, and the USA (South and North America). However, China consumes 88% of the blue water imported through oilseeds domestically while re-exporting ~12%. Among other major commodities, China re-exports 16% of blue VW imported through other grains from the USA (North America) and 10% of blue VW imported through sugar crops from Brazil (South America).

The source countries for green VW trade by China are almost similar to blue VW ones (figure 4).

However, there are considerable differences in quantities. About 44% of green VW embodied in China's oilseeds imports comes from North America and 55% from South America. Out of the 33.4 billion m<sup>3</sup> and 41.9 billion m<sup>3</sup> of green VW imported from North and South America, 3.0 billion m<sup>3</sup> and 4.7 billion m<sup>3</sup> are re-exported by China to third countries. With a total share of 71%, the import of green VW through cotton trade by China is dominated by two regions (South Asia (India) and North America (USA)), of which 43% and 33% of green VW is re-exported to third countries. Although the USA is one of the biggest netexporters of green VW to China through cotton, it imports 11% of the green VW back through the cotton trade.

### 4. Discussion and conclusion

The sustainable use of water and other natural resources is one of the biggest challenges facing humanity. The interaction between domestic and global solutions to resource sustainability has the potential to furnish novel policy measures. Choosing the most appropriate method for tracing VW flows is essential to draw meaningful conclusions about the sustainable use of water resources, as intermediate trade is of growing importance globally (Kanemoto *et al* 2012, Jiang and Guan 2017, Zhang *et al* 2017b, Cadarso *et al* 2018). In this case, we need a decomposition of intermediate product trade between different countries based on backward industrial linkages.

This study used the GVC method to quantify VW trade measures of China's import, export, and re-export of primary and processed agricultural



commodities. Our approach shows the potential to present a more accurate picture of China's role in global VW trade and allocate final VW consumption to final consumers appropriately. The application of the GVC method to the analysis of both blue and green VW trade can be regarded as an initial effort to understand China's role in managing global water resources and the possibility for the international community to collaborate on developing policy measures based on each region's or country's position in the supply and demand for water resources.

Our results indicate that China is a net importer of blue VW through its agricultural commodity trade. This net-import was mainly associated with large cotton imports from South Asian countries (e.g. India and Pakistan). China's significant exports of textile products to North America, Europe, and East Asian countries are associated with large net-exports of blue VW (12.4 billion m<sup>3</sup>). The GVC analysis reveals that  $\sim$ 37% of China's blue VW imported via agricultural commodities is ultimately re-exported to other countries. This finding implies that discussions of sustainable water resource use between China and its trading partners should account for the true complexity of global supply chains-including where water is initially used to produce a good and who realizes the benefits of that good as it travels to the end consumer.

For green VW transfers, China is a substantial net importer (57.9 billion m<sup>3</sup>), mainly through soy imports (72.0 billion m<sup>3</sup>) from the USA, Brazil, and Argentina and cotton (14.1 billion m<sup>3</sup>) from India. Notably, the total blue and green VW imports via soy trade are by far the largest among agricultural commodities for China (75.1 billion m<sup>3</sup>). These results agree with earlier studies that also found that massive imports of soy are transferring vast VW volumes to China (Dalin et al 2014, Ali et al 2017, Hou et al 2018). Out of the total green VW net-imports, we have shown that China re-exports 17% (15.7 billion m<sup>3</sup>) of the imported green VW to the rest of the world. Due to China's increasing role in forward-linkages (i.e. supplier of inputs for other countries' export), the country is expected to export higher volumes of VW through its agricultural trade in the coming years. Therefore, China should seek to utilize its local water resources sustainably while also developing mutually beneficial economic ties that ensure the responsible sourcing of VW for China's imports. While our analysis focused on 2011 (the most recent year for which GVC data were available), future analyses that incorporate more detailed and disaggregated crop and livestock sectors and that assess the temporal evolution of China's VW trade can shed light on the extent to which VW re-export is an increasingly important consideration.

Devising policies for sustainable resource use at national and global levels requires accurate estimation methods. As a large fraction of China's imported VW through agricultural trade is ultimately consumed in other countries, global discussions on the sustainable use of water resources should take into consideration the shared responsibility of nations to act with increasingly complex supply chains in mind. Understanding how VW flows through GVCs and who the intermediate users and end consumers are can facilitate informed decision-making related to water sustainability. Countries with higher resource use efficiencies could be incentivized, for instance, while other countries may be encouraged to conserve their water resources. As China re-exports large proportions of imported VW through agricultural trade, the country can improve its domestic resource use efficiency to supply more resource-efficient products to the world. China (and other major trading countries) should focus on a combination of supply- and demand-side policies and actions for improving the sustainability of global supply chains. Future China-focused research can seek to develop more sophisticated models and more accurate datasets to target policies for specific traded commodities, structural changes of agricultural trade, and their spatial and temporal evolution trends.

## Data availability statement

All data that support the findings of this study are included within the article (and any supplementary files).

#### **ORCID** iDs

Tariq Ali © https://orcid.org/0000-0001-6083-6959 Wei Xie © https://orcid.org/0000-0001-6205-4301 Kyle Frankel Davis © https://orcid.org/0000-0003-4504-1407

#### References

- Ali T, Huang J, Wang J and Xie W 2017 Global footprints of water and land resources through China's food trade *Glob. Food. Security* **12** 139–45
- Allan T 1997 Virtual water: a long-term solution for water-short Middle Eastern economies? *Paper presented at the 1997 British Association Festival of Science* (Leeds: University of Leeds)
- Cadarso M-Á, Monsalve F and Arce G 2018 Emissions burden shifting in global value chains–winners and losers under multi-regional versus bilateral accounting *Econ. Syst. Res.* **30** 439–61
- Chapagain A K and Hoekstra A Y 2004 Water footprints of nations, value of water research Report Series No. 16 (Delft: UNESCO-IHE)
- Chapagain A K, Hoekstra A Y and Savenije H H G 2006 Water saving through international trade of agricultural products *Hydrol. Earth Syst. Sci.* 10 455–68
- D'Odorico P et al 2018 The global food-energy-water nexus Rev. Geophys. 56 456–531
- Dalin C, Hanasaki N, Qiu H, Mauzerall D L and Rodriguez-Iturbe I 2014 Water resources transfers through Chinese interprovincial and foreign food trade *Proc. Natl Acad. Sci. USA* 111 9774–9
- Dalin C, Konar M, Hanasaki N, Rinaldo A and Rodriguez-Iturbe I 2012 Evolution of the global virtual water trade network *Proc. Natl Acad. Sci. USA* **109** 5989–94
- Dalin C, Qiu H, Hanasaki N, Mauzerall D L and Rodriguez-Iturbe I 2015 Balancing water resource conservation and food security in China Proc. Natl Acad. Sci. USA 112 4588
- Davis K F *et al* 2017 Increased food production and reduced water use through optimized crop distribution *Nat. Geosci.* **10** 919–24
- Deng C, Zhang G, Li Z and Li K 2020 Interprovincial food trade and water resources conservation in China Sci. Total Environ. 1 139651

- Elms D K and Low P 2013 *Global Value Chains in a Changing World* (Geneva: World Trade Organization)
- Fader M, Gerten D, Thammer M, Heinke J, Lotze-Campen H, Lucht W and Cramer W 2011 Internal and external green-blue agricultural water footprints of nations, and related water and land savings through trade *Hydrol. Earth Syst. Sci.* 15 1641–60
- FAO 2016 AQUASTAT main database, food and agriculture organization of the United Nations (available at: www.fao.org/nr/water/aquastat/data/query/index. html?lang=en)
- FAO 2021 FAOSTAT online database (available at: www.fao.org/ faostat/en/#data)
- Feng K, Chapagain A, Suh S, Pfister S and Hubacek K 2011 Comparison of bottom-up and top-down approaches to calculating the water footprints of nations *Econ. Syst. Res.* 23 371–85
- Flach R, Ran Y, Godar J, Karlberg L and Suavet C 2016 Towards more spatially explicit assessments of virtual water flows: linking local water use and scarcity to global demand of Brazilian farming commodities *Environ. Res. Lett.* 11 075003
- Greenville J, Kawasaki K and Jouanjean M 2019 Dynamic changes and effects of agro-food GVCS OECD Food, Agriculture and Fisheries Papers, No. 119 (Paris: OECD) (https://www.oecdilibrary.org/agriculture-and-food/dynamic-changes-andeffects-of-agro-food-gvcs\_43b7bcec-en#:~:text=Dynamic% 20Changes%20and%20Effects%20of%20Agro-Food% 20GVCS%20Global,from%20trade%20and% 20the%20impacts%20of%20trade%20policies)
- Grizzetti B, Lanzanova D, Liquete C, Reynaud A and Cardoso A C 2016 Assessing water ecosystem services for water resource management *Environ. Sci. Policy* **61** 194–203
- Hanasaki N 2016 Estimating virtual water contents using a global hydrological model *Terrestrial Water Cycle and Climate Change: Natural and Human-Induced Impacts* ed Q Tang and T Oki (Hoboken, NJ: Wiley) pp 209–28
- Haqiqi I, Taheripour F, Liu J and Van Der Mensbrugghe D 2016 Introducing irrigation water into GTAP database version 9 *J. Glob. Econ. Anal.* **1** 116–55
- Hou S, Liu Y, Zhao X, Tillotson M R, Guo W and Li Y 2018 Blue and green water footprint assessment for china—a multi-region input-output approach *Sustainability* 10 2822
- Jiang X and Guan D 2017 The global CO<sub>2</sub> emissions growth after international crisis and the role of international trade *Energy Policy* **109** 734–46
- Kanemoto K, Lenzen M, Peters G P, Moran D D and Geschke A 2012 Frameworks for comparing emissions associated with production, consumption, and international trade *Environ*. *Sci. Technol.* 46 172–9
- Konar M, Dalin C, Hanasaki N, Rinaldo A and Rodriguez-Iturbe I 2012 Temporal dynamics of blue and green virtual water trade networks *Water Resour. Res.* **48** W07509
- Koopman R, Wang Z and Wei S J 2014 Tracing value-added and double counting in gross exports Amer. Econ. Rev. 104 459–94
- Lin J, Pan D, Davis S J, Zhang Q, He K, Wang C, Streets D G, Wuebbles D J and Guan D 2014 China's international trade and air pollution in the United States *Proc. Natl Acad. USA* 111 1736–41
- Liu J, Zehnder A J B and Yang H 2007 Historical trends in China's virtual water trade *Water Int.* **32** 78–90
- Meier T, Christen O, Semler E, Jahreis G, Voget-Kleschin L, Schrode A and Artmann M 2014 Balancing virtual land imports by a shift in the diet. Using a land balance approach to assess the sustainability of food consumption. Germany as an example *Appetite* 74 20–34
- Mekonnen M M and Hoekstra A Y 2016 Four billion people facing severe water scarcity *Sci. Adv.* **2** e1500323
- Orlowsky B, Hoekstra A Y, Gudmundsson L and Seneviratne S I 2014 Today's virtual water consumption and trade under future water scarcity *Environ. Res. Lett.* **9** 074007

- Qian Y, Tian X, Geng Y, Zhong S, Cui X, Zhang X, Moss D A and Bleischwitz R 2019 Driving factors of agricultural virtual water trade between China and the belt and road countries *Environ. Sci. Technol.* **53** 5877–86
- Salmoral G and Yan X 2018 Food-energy-water nexus: a life cycle analysis on virtual water and embodied energy in food consumption in the Tamar catchment, UK *Resour. Conserv. Recycling* **133** 320–30
- Shi J, Liu J and Pinter L 2014 Recent evolution of China's virtual water trade: analysis of selected crops and considerations for policy *Hydrol. Earth Syst. Sci.* 18 1349–57
- Siebert S and Döll P 2010 Quantifying blue and green virtual water contents in global crop production as well as potential production losses without irrigation *J. Hydrol.* **384** 198–217
- UNESCO 2016 The United Nations world water development report 2016: water and jobs: facts and figures (available at: http://unesdoc.unesco.org/images/0024/002440/ 244041e.pdf)
- Veldkamp T I E *et al* 2017 Water scarcity hotspots travel downstream due to human interventions in the 20th and 21st century *Nat. Commun.* **8** 15697
- Vora N, Shah A, Bilec M M and Khanna V 2017 Food–energy–water nexus: quantifying embodied energy and GHG emissions from irrigation through virtual water transfers in food trade ACS Sustain. Chem. Eng. 5 2119–28
- Vörösmarty C J, Hoekstra A Y, Bunn S E, Conway D and Gupta J 2015 Fresh water goes global *Science* **349** 478–9
- Wang X M, Huang K, Yu Y J, Hu T T and Xu Y J 2016 An input-output structural decomposition analysis of changes in sectoral water footprint in China *Ecol. Indic.* 69 26–34
- Wang Z, Wei S and Zhu K 2013 Quantifying international production sharing at the bilateral and sector levels *NBER Working Paper Series No. 19677* (Cambridge: NBER)

- WTO 2019 Recent patterns of global production and GVC participation Global Value Chain Development Report 2019: Technological Innovation, Supply Chain Trade, and Workers in a Globalized World (Beijing: WTO) (https://doi.org/10.30875/6aa1a271-en)
- Yang Z, Liu H, Xu X and Yang T 2016 Applying the water footprint and dynamic structural decomposition analysis on the growing water use in China during 1997–2007 *Ecol. Indic.* 60 634–43
- Zhang Y, Zhang J, Tang G, Chen M and Wang L 2016 Virtual water flows in the international trade of agricultural products of China *Sci. Total Environ.* 557–558 1–11
- Zhang Z, Zhu K and Hewings G J D 2017a A multi-regional input-output analysis of the pollution haven hypothesis from the perspective of global production fragmentation *Energy Econ.* 64 13–23
- Zhang Z, Zhu K and Hewings G J D 2017b The effects of border-crossing frequencies associated with carbon footprints on border carbon adjustments *Energy Econ.* 65 105–14
- Zhao D, Hubacek K, Feng K, Sun L and Liu J 2019 Explaining virtual water trade: a spatial-temporal analysis of the comparative advantage of land, labor and water in China *Water Res.* **153** 304–14
- Zhao X, Li Y P, Yang H, Liu W F, Tillotson M R, Guan D, Yi Y and Wang H 2018 Measuring scarce water saving from interregional virtual water flows in China *Environ. Res. Lett.* **13** 054012
- Zhao X, Yang H, Yang Z, Chen B and Qin Y 2010 Applying the input-output method to account for WF and VWT in Haihe River Basin in China *Environ. Sci. Technol.* **44** 9150–6
- Zhuo L, Mekonnen M M and Hoekstra A Y 2016 Consumptive water footprint and virtual water trade scenarios for China—with a focus on crop production, consumption and trade *Environ. Int.* 94 211–23